

Analysis of flow differences between the EWQ and post-EWQ periods:



Flow was roughly the same between EWQ and post-EWQ periods. Too few post-EWQ samples were collected (n=14). The range of flow conditions sampled was wider in the EWQ period, and the EWQ data represented the full range of flow conditions better than the post-EWQ data. Not many EWQ samples were collected at high flow, but without that high flow regime represented in post-EWQ data, there is a possibility that water quality differences can falsely interpreted as significant when they really are not. This point is considered in each analysis to follow.



At this location, the Bushkill has not fully mixed with the Delaware – composite sampling is necessary to capture the unequal water quality between the PA and NJ sides of the river.



Annual May to September flow statistics are plotted above. These are flow measurements or estimates associated with the time of each water quality sample. Mean annual flow is about 8,070 cfs; and average May to September flow is about 5,770 cfs. Though DRBC sampled a wide range of flows, these data are most representative of summer flow conditions. Flows were calculated using instantaneous water discharge data from the USGS gage No. 01446500 on the Delaware River at Belvidere times a drainage area ratio.

Upstream ICP: Delaware River at Belvidere 1978 ICP Downstream ICP: Del. River at Riegelsville 1748 ICP

<u>Tributary BCP Watersheds in Upstream Reach:</u> Bushkill Creek, PA – 1841 BCP Delaware River at Sandts Eddy – 1891 MCP* Martins Creek, PA – 1907 BCP Del. River at Martins Creek RR Bridge – 1908 PADEP* Pequest River, NJ – 1978 BCP

*Non-ICP/BCP monitoring locations used to model river water quality. MCP designates "modeling control point"; PADEP designates a site monitored as part of the Pennsylvania Water Quality Network.

Remaining upstream tributaries are less than 20 square miles watershed size, and are not expected to have major water quality effects upon the Delaware River.

Alkalinity as CaCO3, Total mg/l

Existing Water Quality (Table 2I):

Median 34 mg/l Lower 95% Confidence Interval 30 mg/l Upper 95% Confidence Interval 39 mg/l Defined in regulations as a flow-related parameter





No water quality degradation is evident here. Alkalinity apparently did not measurably change between the EWQ and post-EWQ periods.



H0: $\theta_1 = \theta_2 = 0...$ The median of the populations are all equal. H1: $\theta_1 \neq \theta_1$ for at least one i,j The median of the populations are not all equal. ¹ Do not reject the null hypothesis at the 5% significance level.

Sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling frequency, and post-EWQ under-representation of flow conditions. Alkalinity is inversely related to flow in both data sets, though weakly in the EWQ period. Post-EWQ median alkalinity fell within EWQ 95% confidence intervals. Flow is plotted on logarithmic scale. The full flow regime was not fully represented in the post-EWQ data, as noted by the short regression line in the flow vs. concentration plot. There were no independent samples available for comparison with DRBC data.

Ammonia Nitrogen as N, Total mg/l

Existing Water Quality (Table 2I):

Median <0.05 mg/l

Lower 95% Confidence Interval <0.05 mg/l Upper 95% Confidence Interval <0.05 mg/l





No water quality degradation is evident here. Ammonia concentrations apparently declined. Sources of analytical uncertainty included potential laboratory artifacts, detection limit differences, insufficient post-EWQ sampling frequency, and post-EWQ underrepresentation of flow conditions.





¹ Reject the null hypothesis in favour of the alternative hypothesis at the 5% significance level.

Post-EWQ median ammonia concentration was below the EWQ lower 95% confidence interval. We were better able to measure ammonia in the post-EWQ period due to improved detection limits. No independent data were available to validate DRBC results. Flow is plotted on a logarithmic scale. EWQ data possessed 27/40 undetected results that interfered with estimation of the EWQ median. With post-EWQ lower detection levels there were 3/15 undetected results, avoiding such interference and revealing lower concentrations. Some water quality improvement possibly took place, as post-EWQ concentrations were less than 0.04 mg/l. Of course this could also be a laboratory artifact.

Chloride, Total mg/l

Existing Water Quality (Table 2I):

Median 16 mg/l Lower 95% Confidence Interval 14 mg/l Upper 95% Confidence Interval 17 mg/l Defined in regulations as a flow-related parameter





Some evidence of water quality degradation is presented. Median chloride concentrations apparently rose between the two periods. Analytical uncertainty sources included potential laboratory artifacts, insufficient post-EWQ sampling frequency, and flow differences.







Post-EWQ median chlorides rose to the EWQ upper 95% confidence interval. Chloride concentration is related to flow in the EWQ data, but weakly related to flow in the post-EWQ data due to too few post-EWQ samples (n=15). Higher flow conditions were not well-represented in post-EWQ data, so there were fewer low concentrations measured. Flow is plotted on a logarithmic scale. No other data were available to validate DRBC results.

Dissolved Oxygen (DO) mg/l

Existing Water Quality (Table 2I):

Median 8.10 mg/l

Lower 95% Confidence Interval 7.90 mg/l Upper 95% Confidence Interval 8.58 mg/l









No water quality degradation is evident here. No measurable change took place between the EWQ and Post-EWQ periods. Post-EWQ median DO concentration fell within the EWQ 95% confidence intervals. There were too few post-EWQ samples collected (n=14). DO concentration is unrelated to flow in both data sets. Flow is plotted on a logarithmic scale.

Dissolved Oxygen Saturation %

Existing Water Quality (Table 2I):

Median 95%

Lower 95% Confidence Interval 92% Upper 95% Confidence Interval 96%









No water quality degradation is evident here. Dissolved Oxygen Saturation is unrelated to flow, and did not measurably change between the EWQ and post-EWQ periods. Analytical uncertainty sources included insufficient post-EWQ sampling frequency. Post-EWQ median DO saturation fell below the lower EWQ 95% confidence interval, but there were insufficient data to indicate a measurable change. Flow is plotted on a logarithmic scale. There were two low saturation values of 73% and 77% found in July 2010 and August 2011, respectively under low flow conditions. Biweekly instead of monthly sampling is recommended for this location. No independent data were available for comparison with DRBC results.

Enterococcus colonies/100 ml

Existing Water Quality (Table 2I):

Median 145/100 ml Lower 95% Confidence Interval 80/100 ml Upper 95% Confidence Interval 250/100 ml









No water quality degradation is evident here. Enterococci apparently declined between the EWQ and Post-EWQ periods. Analytical uncertainty sources included potential laboratory artifacts, insufficient post-EWQ sampling frequency, and flow differences. Enterococcus concentrations are unrelated to flow in both data sets. Concentrations and flows are plotted on logarithmic scale, and regressions are power relationships. Post-EWQ median enterococcus concentrations fell below the lower EWQ 95% confidence interval. No independent data were available for comparison with DRBC results at this location.

Escherichia coli colonies/100 ml

Existing Water Quality (Table 2I):

Median 31/100 ml

Lower 95% Confidence Interval 24/100 ml Upper 95% Confidence Interval 64/100 ml Defined in regulations as a flow-related parameter





No water quality degradation is evident here. E. coli concentrations apparently did not measurably change between the EWQ and Post-EWQ periods. Analytical uncertainty sources included potential laboratory artifacts, insufficient post-EWQ sampling frequency, and flow differences.





Post-EWQ median E. coli fell within the EWQ 95% confidence intervals. Concentrations and flows are plotted on logarithmic scale, and regressions are power relationships. E. coli concentrations are very weakly related to flow in the EWQ data set, but positively related to flow in the post-EWQ data set. No independent data were available to validate DRBC results at this location.

Fecal coliform colonies/100 ml

Existing Water Quality (Table 2I):

Median 100/100 ml Lower 95% Confidence Interval 64/100 ml Upper 95% Confidence Interval 130/100 ml









No water quality degradation is evident here. Fecal coliform concentrations apparently fell below the lower EWQ 95% confidence interval, but did not measurably change between the EWQ and post-EWQ periods. Analytical uncertainty sources included laboratory artifacts, insufficient post-EWQ sampling frequency (n=15), and flow differences. Fecal coliform is unrelated to flow in both data sets. Post-EWQ data were not fully representative of the flow regime. Concentrations and flows are plotted on logarithmic scale. No independent data were available for comparison with DRBC data.

Hardness as CaCO3, Total mg/l

Existing Water Quality (Table 2I):

Median 48 mg/l

Lower 95% Confidence Interval 45 mg/l Upper 95% Confidence Interval 52 mg/l









No water quality degradation is evident here. Hardness apparently did not measurably change between the EWQ and post-EWQ periods. Analytical uncertainty sources included potential laboratory artifacts, insufficient post-EWQ sampling frequency, and flow differences. Hardness is weakly related to flow in both data sets, though the strength of the relationship in the post-EWQ data is influenced by a single outlier value. Post-EWQ median hardness fell within the EWQ 95% confidence intervals. Flow is plotted on a logarithmic scale. No independent data were available for comparison with DRBC results.

Nitrate + Nitrite as N, Total mg/l

Existing Water Quality (Table 2I, as Nitrate only):

Median 0.85 mg/l

Lower 95% Confidence Interval 0.70 mg/l Upper 95% Confidence Interval 0.90 mg/l





No water quality degradation is evident here. Nitrate concentrations apparently declined between the EWQ and post-EWQ periods. Analytical uncertainty sources included potential laboratory artifacts, detection limit differences, insufficient post-EWQ sampling frequency, and flow differences.





Nitrate is unrelated to flow in both data sets. Post-EWQ nitrate concentrations fell below the lower EWQ 95% confidence interval. Post-EWQ nitrate + nitrite concentrations were assumed equivalent for comparison with EWQ nitrate concentrations since EWQ nitrite concentrations were never detected. Independent data were not available for validation of DRBC data.

Nitrogen as N, Total (TN) mg/l

Existing Water Quality (Table 2I):

Median 1.19 mg/l

Lower 95% Confidence Interval 1.01 mg/l Upper 95% Confidence Interval 1.35 mg/l







No water quality degradation is evident here. Total Nitrogen concentrations apparently declined between the EWQ and post-EWQ periods. Analytical uncertainty sources included potential laboratory artifacts, detection limit differences, insufficient post-EWQ sampling frequency, and flow differences. TN is unrelated to flow in both data sets. Post-EWQ median TN concentrations fell below the EWQ lower 95% confidence interval. No independent data were available to validate DRBC results.

Nitrogen, Kjeldahl as N, Total (TKN) mg/l

Existing Water Quality (Table 2I):

Median 0.35 mg/l

Lower 95% Confidence Interval 0.26 mg/l Upper 95% Confidence Interval 0.46 mg/l







No water quality degradation is evident here. TKN concentrations apparently declined between the EWQ and post-EWQ periods. However, analytical uncertainty sources included potential laboratory artifacts, insufficient post-EWQ sampling frequency, and flow differences. The post-EWQ range was far narrower and all concentrations were less than 0.33 mg/l. TKN concentration is unrelated to flow in both data sets. Post-EWQ median TKN fell below the lower EWQ 95% confidence interval. There were no independent data to confirm DRBC results.

Orthophosphate as P, Total mg/l (OP)

Existing Water Quality (Table 2I):

Median 0.02 mg/l

Lower 95% Confidence Interval 0.01 mg/l Upper 95% Confidence Interval 0.02 mg/l





No water quality degradation is evident here. OP concentrations apparently declined between the EWQ and post-EWQ periods. However, analytical uncertainty sources included potential laboratory artifacts, detection limit differences, insufficient post-EWQ sampling frequency, and flow differences.



OP is unrelated to flow in both data sets. Post-EWQ median orthophosphate fell to the EWQ lower 95% confidence interval. This may indicate a water quality improvement in that there were no post-EWQ concentrations higher than 0.02 mg/l, but could also be a laboratory artifact. The EWQ orthophosphate undetected results rate was 16 out of 43 samples. The detection limit was lower in the post-EWQ results, and there were no undetected results. Post-EWQ orthophosphate ranged narrower than EWQ data. There were no independent data to confirm DRBC results.

рΗ

Existing Water Quality (Table 2I):

Median 7.55 standard units

Lower 95% Confidence Interval 7.41 standard units Upper 95% Confidence Interval 7.70 standard units





No water quality degradation is evident here. pH apparently did not measurably change between the EWQ and post-EWQ periods. However, analytical uncertainty sources included insufficient post-EWQ sampling frequency and flow differences.





pH is weakly related to flow in the EWQ data set and inversely related to flow in the post-EWQ data. pH tends toward neutral at higher flow conditions. Post-EWQ median pH was near the upper EWQ 95% confidence interval, but there were too few data (n=14) to measure a true difference. In July 2010 there was one spike above pH 9, indicating high algal productivity during that dry sampling period. There is a very large aquatic plant bed just upstream of this location. No independent data were available to confirm DRBC results.

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Existing Water Quality (Table 2I):

Median 0.05 mg/l Lower 95% Confidence Interval 0.04 mg/l

Upper 95% Confidence Interval 0.06 mg/l







No water quality degradation is evident here. Total Phosphorus (TP) concentrations apparently declined between the EWQ and post-EWQ periods. However, analytical uncertainty sources included potential laboratory artifacts, detection limit differences, insufficient post-EWQ sampling frequency, and flow differences. Post-EWQ median total phosphorus fell below the EWQ lower 95% confidence interval. TP is unrelated to flow in both data sets. No independent data were available to confirm DRBC results.

Specific Conductance µmho/cm

Existing Water Quality (Table 2I):

Median 142 µmho/cm

Lower 95% Confidence Interval 127 µmho/cm Upper 95% Confidence Interval 155 µmho/cm Defined in regulations as a flow-related parameter





No water quality degradation is evident here. Specific conductance apparently did not measurably change between the EWQ and post-EWQ periods. However, analytical uncertainty sources included insufficient post-EWQ sampling frequency and flow differences.







Post-EWQ median specific conductance fell within the EWQ 95% confidence intervals. Specific conductance is inversely related to flow in both data sets. The rise in specific conductance seen elsewhere is not apparent here; probably because the full flow regime is not well-represented in the post-EWQ data. Median specific conductance rose by 10 μ mho/cm but there were insufficient post-EWQ samples (n=14) for a valid comparison. No independent data were available for comparison with DRBC results.

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Total Dissolved Solids (TDS) mg/l

Existing Water Quality (Table 2I):

Median 110 mg/l Lower 95% Confidence Interval 103 mg/l Upper 95% Confidence Interval 120 mg/l





No water quality degradation is evident here. TDS apparently declined between the EWQ and post-EWQ periods. However, analytical uncertainty sources included potential laboratory artifacts, detection limit differences, insufficient post-EWQ sampling frequency, and flow differences. TDS is unrelated to flow in both data sets.



Post-EWQ median TDS fell below the EWQ lower 95% lower confidence interval, and were less variable than the baseline samples as well. Post-EWQ detection limits were lower than EWQ detection limits, though there were no non-detect results at any time. Perhaps this decline is not real but an artifact of different laboratory practices. The annual plot shows clear differences in data from three different contract labs used in 2000, 2001-2004 and 2009-2011. To solve or at least account for this problem, blind standards and split samples should be sent to different laboratories used by the monitoring program.

Total Suspended Solids (TSS) mg/l

Existing Water Quality (Table 2I):

Median 4.0 mg/l Lower 95% Confidence Interval 3.0 mg/l Upper 95% Confidence Interval 5.0 mg/l Defined in regulations as a flow-related parameter









ull hypothesis at the 5% significance level

No water quality degradation is evident here. TSS apparently did not measurably change between the EWQ and post-EWQ periods. However, analytical uncertainty sources included potential laboratory artifacts, insufficient post-EWQ sampling frequency, and flow differences. TSS is positively related to flow in both data sets. Post-EWQ median TSS fell below the lower EWQ 95% confidence interval, but there were insufficient post-EWQ data (n=14) for valid comparison. Flows and concentrations are plotted on logarithmic scale, and regressions are power relationships. No independent data were available for comparison with DRBC results.

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Turbidity NTU

Existing Water Quality (Table 2I):

Median 2.6 NTU Lower 95% Confidence Interval 1.8 NTU Upper 95% Confidence Interval 4.0 NTU Defined in regulations as a flow-related parameter







No water quality degradation is evident here. Turbidity apparently declined between the EWQ and post-EWQ periods. However, analytical uncertainty sources included insufficient post-EWQ sampling frequency, and flow differences. Post-EWQ median turbidity fell below the lower EWQ 95% confidence interval of the median, but the comparison is not valid due to too few post-EWQ samples (n=14) and no post-EWQ high-flow samples. Turbidity is positively related to flow in both data sets, power regression lines are shown. There were no independent data available for comparison with DRBC results. Water Temperature, degrees C

Not included in DRBC Existing Water Quality rules









No water quality degradation is evident here. Water temperature apparently did not measurably change between the EWQ and post-EWQ periods. However, analytical uncertainty sources included insufficient post-EWQ sampling frequency, and flow differences. Water temperature is inversely related to flow in the EWQ data set, but unrelated to flow in the post-EWQ data. There were insufficient samples taken in the post-EWQ period to account for a sufficient range of flow conditions, thus the lack of relationship of temperature to flow. Note that flows are plotted on a logarithmic scale. No independent data were available for comparison with DRBC results.